

Degradation of PFAS in historically contaminated soils: evidence of biotic processes

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Content of presentation

Context

Why study PFAS biodegradation in contaminated soils?

Materials and methods

Soil characterization, evidence of biodegradation and long-term incubation

Results : short-term evolution

Effect of bacterial inhibitor

Results : long-term evolution

Temporal changes in PFAS concentrations and evidence of precursor transformation

Mechanistic interpretation and perspectives

Working hypotheses and next analytical priorities

Why study PFAS biodegradation in contaminated soils?

1- Major PFAS sources

PFAS
manufacturing
sites



Use of AFFF



Landfill

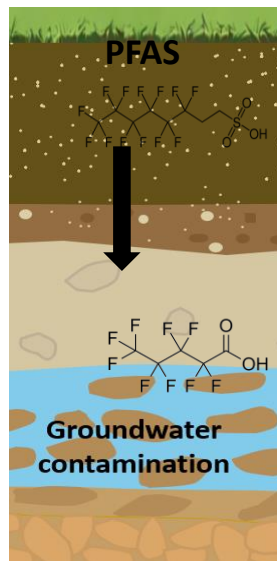


Contaminated
sludge



2- Environmental transfer

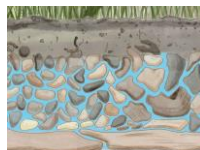
Transfer of PFAS from soil
to groundwater



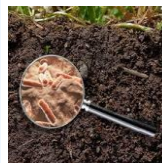
3- Why Soils matter



Long term
reservoir of
PFAS



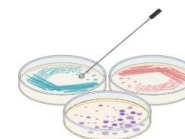
Potential source
of groundwater
contamination



Complex
interactions with
soil constituents
and microorganisms

4- Scientific gap and objective

Most
biodegradation
studies :
Isolated bacterial
strains/
communities



No studies:
Historically
contaminated
soil with
indigenous
microorganisms



Objective:

Understand PFAS biodegradation
pathways in AFFF-contaminated soil:

- Parent compounds
- Degradation products
- Kinetics
- Half-lives

Experimental strategy and soil used

1- **Short time experiment \pm bacterial inhibitor (NaN_3)** to evaluate the possibility of PFAS biodegradation in historically contaminated soil

2- Long term biodegradation experiment to further investigate precursor transformation and degradation products

Soil used

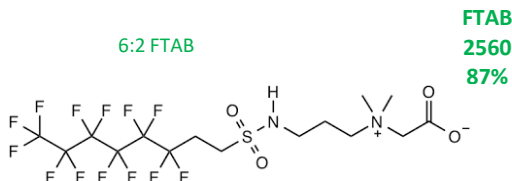
Soil sampled from a historical fire training zone

Physico-Chemical Analysis of the Soil Used

Soil predominantly sandy (78 % sand, 15 % silt, 7 % clay)

Slightly alkaline pH (pH = 8.4)

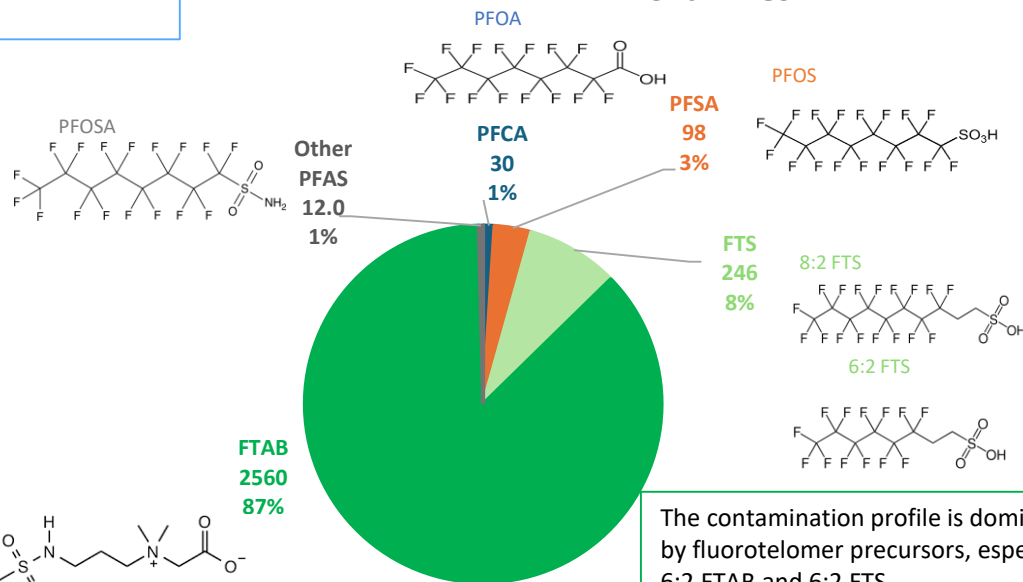
Relatively low Total organic carbon
(TOC = 10 g.kg⁻¹)



PFAS concentration in the soil

Total PFAS concentration = 2,945 $\mu\text{g.kg}^{-1}$ dw

Concentrations (in $\mu\text{g.kg}^{-1}$ dw) of the different PEAS families

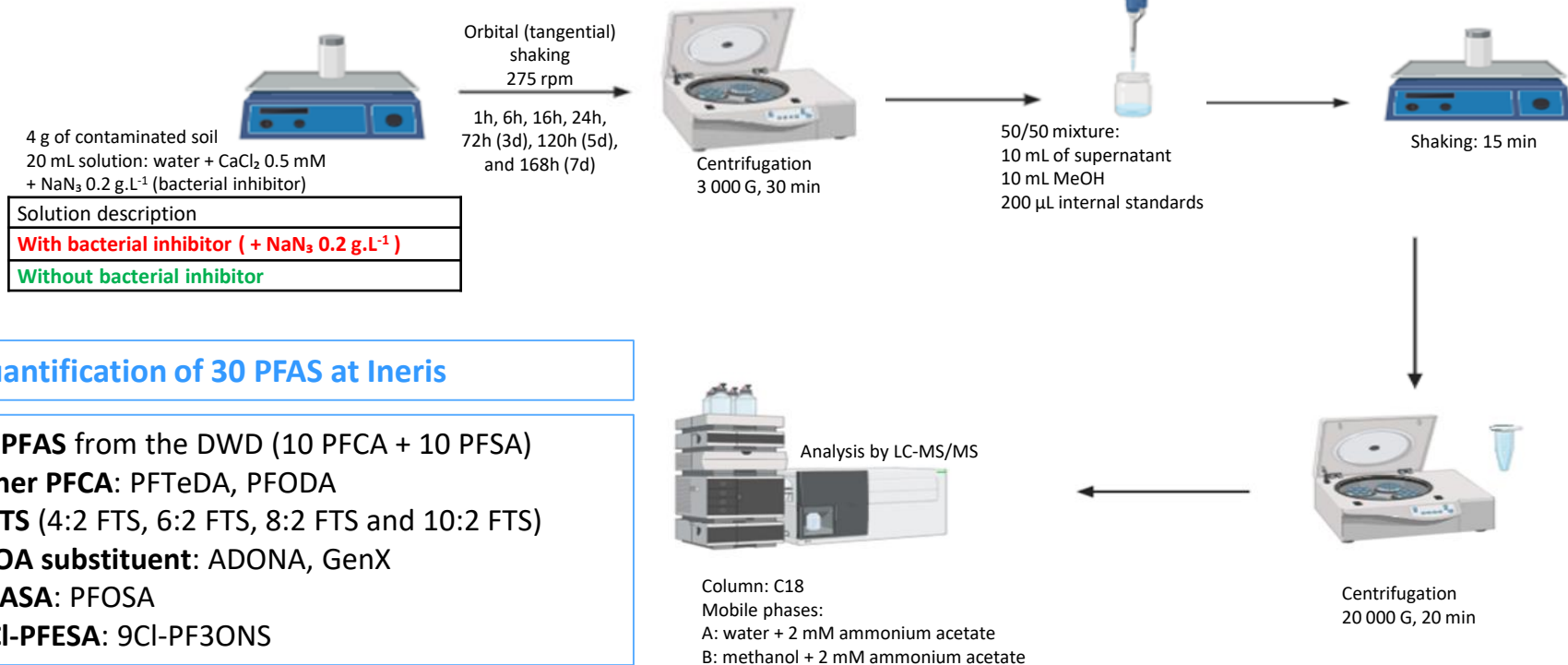


The contamination profile is dominated by fluorotelomer precursors, especially 6:2 FTAB and 6:2 FTS

Materials and methods

1- Short time experiment ± bacterial inhibitor (NaN_3)

Objective: evaluate the possibility of PFAS biodegradation in historically contaminated soil



2- Long-term biodegradation experiment

Experimental setup



80 g contaminated soil
+ 400 mL solution (same
ratio "volume/mass" used in
the first experiment)

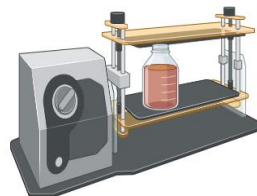
0.5 mM CaCl_2

Two tested conditions

With bacterial inhibitor
(NaN_3 0.2 g.L⁻¹)

**Without bacterial
inhibitor**

Equilibrium



Shaking (3 days)
Reach equilibrium
Separate release
from degradation

Long-term monitoring

Static incubation for 1
year



Sampling over time: 3 mL
collected for each sample

3d 7d 21d — 1 year

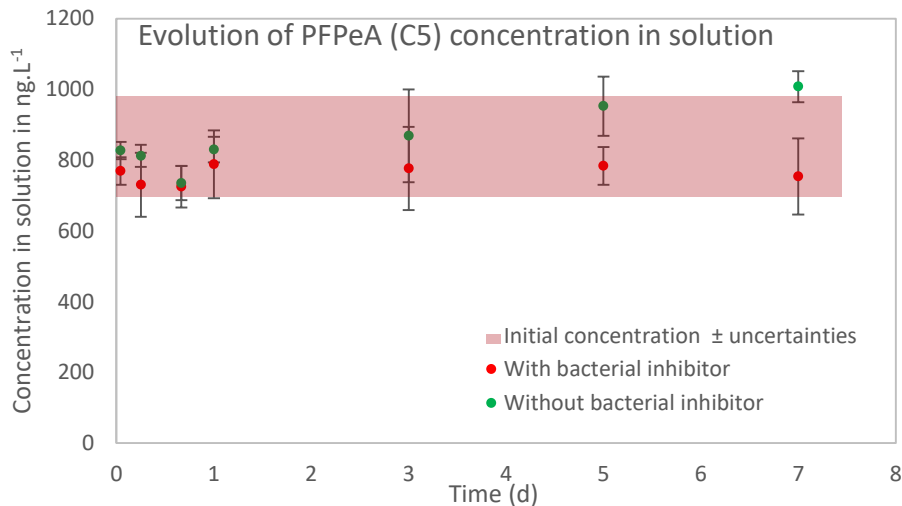
Objectives

- ✓ Confirm whether biodegradation occurs
- ✓ Determine the transformation products formed
- ✓ Identify the parent compounds that degrade
- ✓ Determine biodegradation kinetics and half-lives

Results and discussion

Evaluation of PFAS biodegradation potential in historically contaminated soil

Effect of bacterial inhibitor NaN_3



With bacterial inhibitor

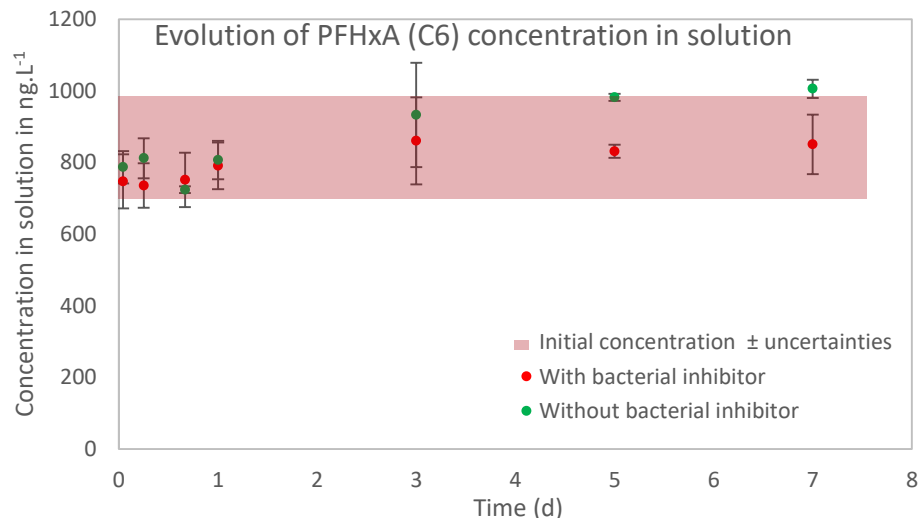
Steady concentration over time

Equilibrium reached rapidly, within 1 hour

Without bacterial inhibitor

- Concentration increased over time
- Concentration significantly higher than with NaN_3 (5–7 days)

Short chain PFCA (PFBA, PFPeA, PFHxA, PFHpA)

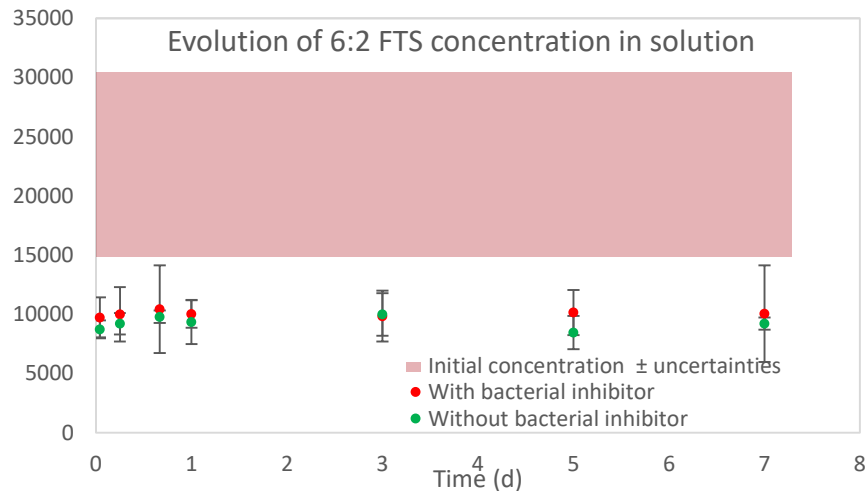


→ Possible biotransformation of precursors (e.g., FTS and / or FTAB)?

Results and discussion

Evaluation of PFAS biodegradation potential in historically contaminated soil

Effect of bacterial inhibitor NaN_3



With bacterial inhibitor Constant concentration over time

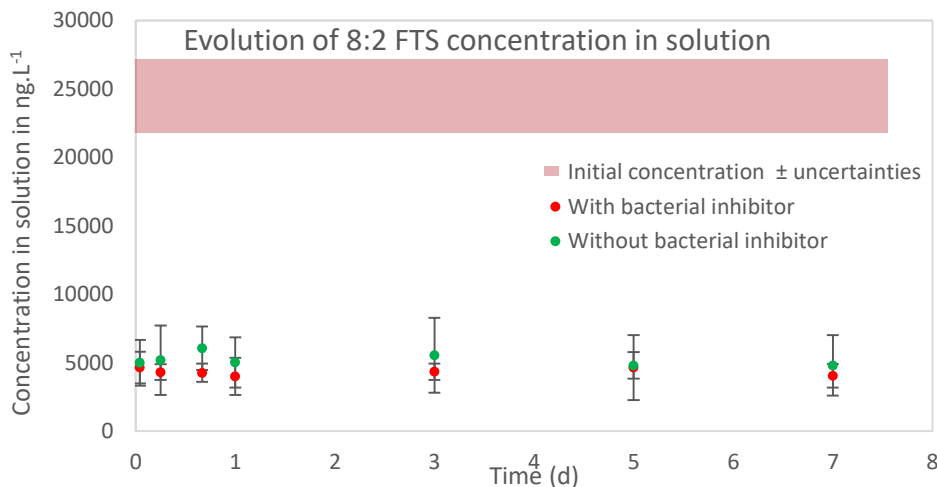
Without bacterial inhibitor

No decrease in FTS concentrations over time as expected

But high concentration in soil

6:2 FTS concentration may therefore be continuously formed while being simultaneously transformed

FTS



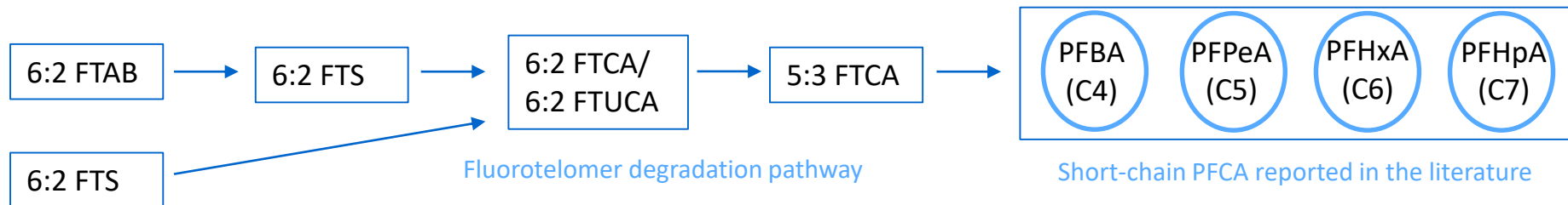
Equilibrium reached rapidly, within 1 hour

- ➔ Possible continuous formation of 6:2 FTS from 6:2 FTAB
- ➔ This may mask 6:2 FTS depletion and maintain apparent equilibrium in solution

Results and discussion

Literature evidence for precursor biodegradation into short-chain PFCA

Possible formation of C4–C7 from precursor degradation : 6:2 FTAB/ 6:2 FTS ?



Adapted from Wang et al. (2011), D'Agostino & Mabury et al. (2017), Zhang et al.(2016), Shaw et al. (2019), Mendez et al. (2022) and Yan et al. (2024)

Evidence for 6:2 FTS degradation

- 1- Wang et al. (2011) : activated sludge from WWTP, mixed microbial community, aerobic conditions and formation of PFPeA (C5) and PFHxA (C6)
- 2- Zhang et al.(2016) : river sediment with natural microbial communities, formation of PFPeA (C5) and PFHxA (C6)
- 3- Shaw et al. (2019): pure culture of *Gordonia* sp. NB4-1Y, rapid transformation of 6:2 FTS with PFBA (C4), PFPeA (C5) and PFHxA (C6) formation
- 4- Mendez et al. (2022): pure culture of *Dietzia aurantiaca* J3 with formation of PFPeA (C5) and PFHxA (C6)
- 5- Yan et al. (2024) : AFFF-contaminated soil microcosms , formation of PFBA (C4), PFPeA (C5), PFHxA (C6) ; PFHpA (C7)

Evidence for 6:2 FTAB degradation

- 1- Shaw et al. (2019): pure culture of *Gordonia* sp. NB4-1Y, 6:2 FTAB biotransformed to PFBA (C4), PFPeA (C5) and PFHxA (C6)
 - 2- D'Agostino & Mabury (2017): aerobic WWTP sludge, slow 6:2 FTAB biodegradation with formation of PFPeA (C5) and PFHxA (C6)
- In both studies, 6:2 FTS may be involved as an intermediate, but it was not clearly detected during 6:2 FTAB biotransformation.*

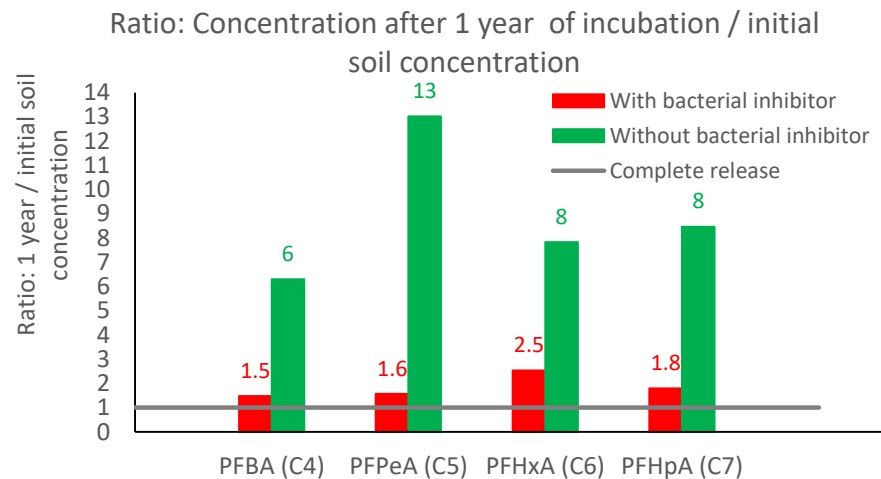
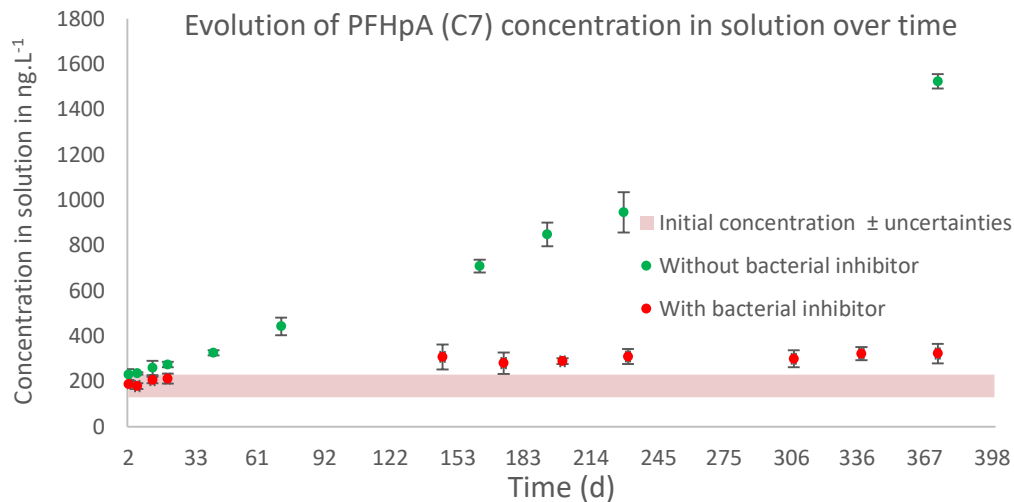
Interpretation for this study

Fluorotelomer precursors can product C4–C7 PFCA

Results and discussion

Long-term result : increase of short-chain PFCA concentration in solution

Example shown for PFHpA (C7); same trend for PFBA (C4), PFPeA (C5) and PFHxA (C6)



With bacterial inhibitor : PFBA (C4), PFPeA (C5), PFHxA (C6) and PFHpA (C7) concentrations in solution increased up to ~150 days and then stabilized

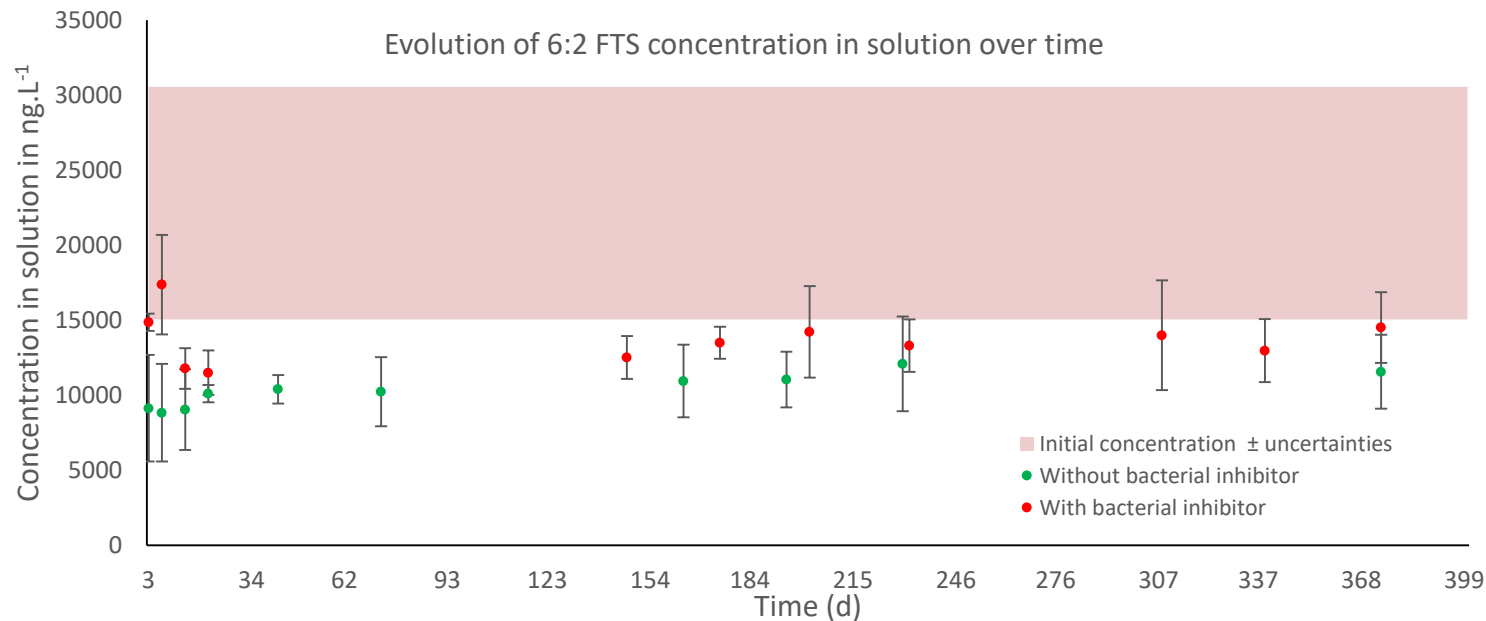
Without bacterial inhibitor : short-chain PFCA concentrations in solution increased much more strongly after 1 year, exceeding the initial concentration

→ NaN_3 inhibition may decrease over long-term incubation
 → Microorganisms may not be fully inhibited

→ Biotransformation of PFAS precursors

Results and discussion

FTS do not show the expected depletion despite PFCA formation



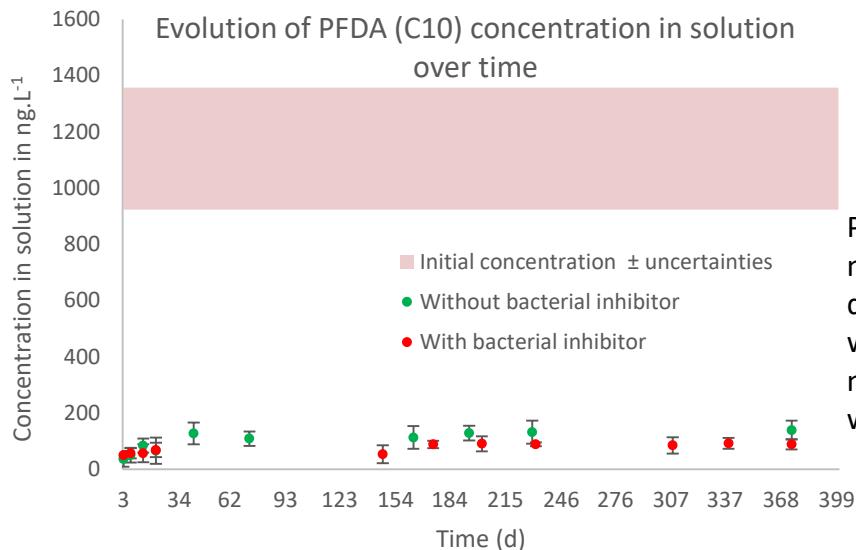
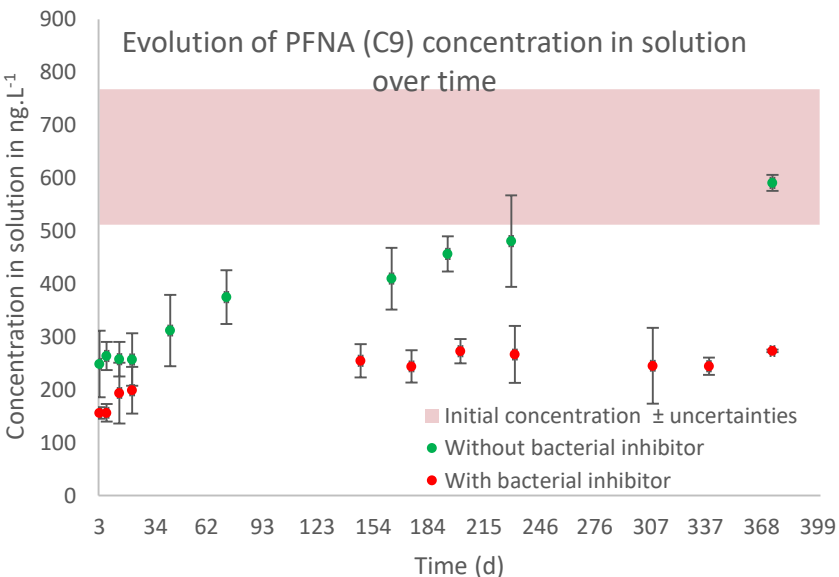
6: 2 FTS (same as 8:2 FTS) remains stable and close to its initial level in both conditions



✓ A Stable 6:2 FTS concentration does not rule out biodegradation :
6:2 FTS may be continuously formed from 6:2 FTAB while being simultaneously biotransformed via 6:2 FTCA/6:2 FTUCA and 5:3 FTCA into short-chain PFCA

Results and discussion

Long-chain PFCA : additional release or biological contribution?



PFUnDA (C11) is near LOQ (limit of quantification) with inhibitor but not detected without inhibitor

PFNA (C9) and PFDA (C10) concentrations in solution increased more markedly in the absence of bacterial inhibitor



The increase in PFNA (C9) and PFDA (C10) without bacterial inhibitor may indicate a biological contribution, possibly through transformation of unidentified long-chain PFCA precursors; however, direct biodegradation of PFUnDA (C11) into PFDA (C10) and PFNA (C9) is not clearly demonstrated in soil systems

Conclusions and perspectives

Summary of the main results

Evidence of PFAS biodegradation (0 –7 d)

○ With bacterial inhibitor (NaN_3) :

- Rapid equilibrium, reached within ~ 1 h
- Concentrations remained globally stable over time

○ Without bacterial inhibitor (NaN_3) :

- Higher concentrations of short-chain PFCA at 5 – 7 d

➔ Short term result : first indicator of biodegradation

Long-term incubation (1 year)

- PFBA (C4), PFPeA (C5), PFHxA (C6) and PFHpA (C7) concentrations in solution increased over time
- After ~ 150 d, concentrations tended to stabilize with inhibitor, but continued to rise much more strongly without inhibitor
- Short-chain PFCA (C4 –C7) increased while monitored FTS concentrations remained globally stable
- PFNA (C9) and PFDA (C10) concentrations increased more markedly in the absence of bacterial inhibitor

➔ Long-term result: concentration increases are consistent with biological contribution in addition to release

A stable 6:2 FTS concentration does not rule out biodegradation

Literature supports that precursor biotransformation can generate short-chain PFCA

Overall, the data support precursor transformation in contaminated soil, in addition to release

Conclusions and perspectives

Mechanistic interpretation : two hypotheses to test next

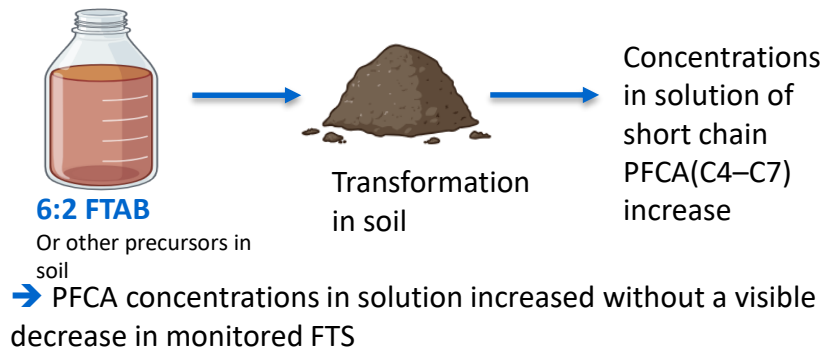
The results point to precursor transformation, but the parent compounds still need to be identified

How can FTS concentrations in solution remain stable while those of short-chain PFCA increased?

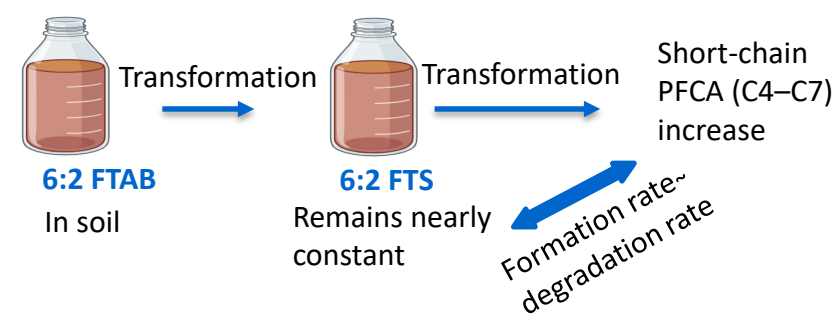
Short-chain PFCA (C4– C7):
Concentration in solution increased over time

6: 2 FTS :
Concentration in solution remained relatively stable

Hypothesis 1 : direct precursor transformation



Hypothesis 2 : Pseudo-steady state of the 6:2 FTS



→ Analysis in progress : analyze the precursor compound 6:2 FTAB to distinguish between the two metabolic pathways

→ Next step: determinate the half-lives and degradation kinetics

Thank you for your attention

References

- D'Agostino, Lisa A., et Scott A. Mabury. 2017. « Aerobic Biodegradation of 2 Fluorotelomer Sulfonamide–Based Aqueous Film–Forming Foam Components Produces Perfluoroalkyl Carboxylates ». *Environmental Toxicology and Chemistry* 36 (8): 2012-21. <https://doi.org/10.1002/etc.3750>.
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- Shaw, Dayton M. J., Gabriel Munoz, Eric M. Bottos, et al. 2019. « Degradation and defluorination of 6:2 fluorotelomer sulfonamidoalkyl betaine and 6:2 fluorotelomer sulfonate by *Gordonia* sp. strain NB4-1Y under sulfur-limiting conditions ». *Science of The Total Environment* 647 (janvier): 690-98. <https://doi.org/10.1016/j.scitotenv.2018.08.012>.
- Wang, Ning, Jinxia Liu, Robert C. Buck, et al. 2011. « 6:2 Fluorotelomer sulfonate aerobic biotransformation in activated sludge of waste water treatment plants ». *Chemosphere* 82 (6): 853-58. <https://doi.org/10.1016/j.chemosphere.2010.11.003>.
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